Zeitschrift: Bulletin of the Geobotanical Institute ETH

Herausgeber: Geobotanisches Institut, ETH Zürich, Stiftung Rübel

Band: 64 (1998)

Artikel: Does moving in summer reduce the abundance of common reed

(Phragmites australis)?

Autor: Güsewell, Sabine

DOI: https://doi.org/10.5169/seals-377814

Nutzungsbedingungen

Die ETH-Bibliothek ist die Anbieterin der digitalisierten Zeitschriften. Sie besitzt keine Urheberrechte an den Zeitschriften und ist nicht verantwortlich für deren Inhalte. Die Rechte liegen in der Regel bei den Herausgebern beziehungsweise den externen Rechteinhabern. Siehe Rechtliche Hinweise.

Conditions d'utilisation

L'ETH Library est le fournisseur des revues numérisées. Elle ne détient aucun droit d'auteur sur les revues et n'est pas responsable de leur contenu. En règle générale, les droits sont détenus par les éditeurs ou les détenteurs de droits externes. <u>Voir Informations légales.</u>

Terms of use

The ETH Library is the provider of the digitised journals. It does not own any copyrights to the journals and is not responsible for their content. The rights usually lie with the publishers or the external rights holders. See Legal notice.

Download PDF: 04.05.2025

ETH-Bibliothek Zürich, E-Periodica, https://www.e-periodica.ch

Does moving in summer reduce the abundance of common reed (*Phragmites australis*)?

Sabine Güsewell

Geobotanisches Institut ETH, Stiftung Rübel, Zürichbergstrasse 38, 8044 Zürich, Switzerland; guesewell@geobot.umnw.ethz.ch

Summary

- 1 Common reed (*Phragmites australis* Trin.) has recently spread in fen meadows of the Swiss Plateau, where it might reduce species richness and displace typical fen species.
- 2 Mowing experiments were carried out in two fens near Zurich to investigate whether mowing in June (in addition to the usual September cutting) is an effective measure to reduce the abundance of *Phragmites*. Changes in the number and size of *Phragmites* shoots were monitored during three years. The aboveground biomass and nutrient concentrations of *Phragmites* and of all other vascular plant species ("other species") were determined in the third year of the experiment.
- 3 The additional June cutting had no significant effect on shoot number and size, and, therefore, on the aboveground biomass of *Phragmites* during this period. The biomass of *Phragmites* did not differ between treatments, but the biomass of the other species was lower in plots with additional June cutting. Due to this additional cutting, 90% more N and 181% more P were, on an average, removed with *Phragmites*, but only 30% more N and 64% more P with the other species. Thus, the nutrient economy of *Phragmites* was stronger affected than the nutrient economy of the rest of the species.
- 4 The availability of nutrients and interspecific competition are probably decisive for the long-term treatment effects. Further monitoring is needed to evaluate whether mowing in June and September will eventually reduce the abundance of *Phragmites* by depleting its belowground reserves. However, the results of this study indicate that other means (e.g. grazing) are necessary for a short-term control.

Keywords: cutting experiment, cutting regime, management, nature conservation, nutrient economy, wetlands

Bulletin of the Geobotanical Institute ETH (1998), 64, 23–35

Introduction

Common reed (*Phragmites australis* Trin.) is a world-wide distributed tall grass species of great economic and ecological importance (Haslam 1973a,b; Rodewald-Rudescu 1974). Many reed stands are managed to increase their production and to regulate the size of

shoots (Haslam 1973a; Granéli 1984), to enhance their mechanical resistance on lake shores (Klötzli 1974; Ostendorp 1995), or to maintain favourable habitat conditions for reed-dwelling birds and insects (Bibby & Lunn 1982; Tscharntke 1992). On the other

hand, Phragmites can be a troublesome weed invading various cultures, e.g. rice or cane (Izatt 1979), fishponds and channels (Husák 1978), coastal areas (Hellings & Gallagher 1992), or freshwater wetlands such as fens (Biewer 1994). Many fen meadows of the Swiss Plateau have recently been affected by a spread of Phragmites (Klötzli 1986; Marti & Müller 1993), which is considered undesirable as it is associated with a decrease in species richness and the disappearance of typical fen species (Güsewell & Klötzli 1998). Even though it is not yet clear to what extent Phragmites is actually responsible for these changes (Güsewell & Klötzli 1998), managers of wetlands nature reserves often seek to control Phragmites at sites where its abundance has increased (Egloff 1984; Bressous et al. 1992).

Different techniques have been applied to control Phragmites. Good results are normally obtained by spraying herbicides (Glyphosate or Dalapon) followed by mowing or burning (Izatt 1979; Jones & Lehman 1987). Mowing or burning followed by deep flooding is another option (Husák 1978; Demina 1979; Hellings & Gallagher 1992). Such "drastic" measures are suitable in species-poor vegetation types strongly dominated by Phragmites, but not in fen meadows, where other species have to be preserved. To control Phragmites at such sites, measures based on the traditional management, to which the protected species are adapted, would be more adequate. In Switzerland, lowland fen meadows have traditionally been mown in late summer or autumn (Egloff 1984). It is generally assumed that mowing earlier in summer (June-July) will reduce the abundance of Phragmites (Egloff 1984; Bressous et al. 1992). Indeed, in fens or wet grasslands that have been regularly mown in summer for a long time little or no Phragmites occurs (van Diggelen et al. 1996). Yet, management experiments provided little direct evidence for the effectiveness of this treatment so far. Phragmites was either not present at the experimental sites (Wolf et al. 1984; Bakker & de Vries 1985; Kapfer & Pfadenhauer 1986; Oomes & Altena 1987; Rosenthal 1992), or changes in its abundance were not related to treatments (Finckh 1960; Rowell et al. 1985; Egloff 1986).

To be an effective control of Phragmites, mowing in June (in addition to September) should reduce its aboveground biomass through a decrease in the number or size of shoots. Moreover, the aboveground biomass of Phragmites should be reduced more strongly than that of the other species present at the site, so that the dominance of Phragmites decreases. To produce such an effect, the additional mowing in early summer should affect the nutrient economy of Phragmites more than that of the species to be preserved. To determine whether additional mowing in early summer actually produces these effects, experiments have been carried out in two fen meadows of the region of Zurich since 1995. This contribution presents results after the first three years of management.

Methods

STUDY SITES AND EXPERIMENTAL DESIGN

The two experimental sites are fens located on the Swiss Plateau near Zurich, at an altitude of 430–440 m a.s.l. The long-term average annual temperature of the area is 8–10 °C, the average annual rainfall 1000–1100 mm. Soils are calcareous humic gleysols, with strongly decomposed and humified peat in the top soil ("Anmoor"). Due to fluctuations of the groundwater table, soils are waterlogged in winter, but relatively dry in summer.

Site "Greifensee" (Swiss National Grid 692'550/247'750) is a wet meadow near lake Greifensee, dominated by Molinia coerulea and various Carex species; Phragmites is moderately abundant (nomenclature follows Hess et al. 1991). A ditch draining nutrientrich water from the adjacent farmland runs through the site. Along this ditch the vegetation is highly productive, dominated by Phragmites, Carex acutiformis, Holcus lanatus and various tall forbs. Site "Katzensee" (680'550/254'100) is situated in the "Allmend Katzensee", a flat swampy basin. The vegetation is dominated by Molinia coerulea, Carex acutiformis and various tall herbs in the drier parts, and by Carex panicea, Carex elata, Juncus subnodulosus and tall herbs in the wetter parts. Both sites had been mown yearly in September for at least five years before the experiments started.

The experiment followed a block design with blocks 1-3 at site "Greifensee" and blocks 4-5 at site "Katzensee". Each block consisted of two plots (10 x 10 m²). Both plots were mown yearly in early September and the litter was removed soon after mowing. The treated plots were additionally mown in late June. Treatments started in 1995, except for block 3, where they only started in 1996. To monitor the effects of treatments on the abundance of *Phragmites*, twelve 1-m² permanent quadrats were established per experimental plot. Quadrats were arranged systematically in three groups of four quadrats (i.e. three 4-m² quadrats) to minimize trampling and edge effects.

FIELD AND LABORATORY TECHNIQUES

The abundance of *Phragmites* in all plots was recorded in late June and late August or early September 1995, 1996 and 1997. The late summer measurements will hereafter be referred to as "September" measurements. The

number of shoots taller than 20 cm was counted in all twelve 1-m² quadrats per plot ("shoot density"). The culm length and the basal diameter of all shoots were measured in one randomly chosen quarter of each 4-m² quadrat. Culm length was measured from the soil surface to the base of the uppermost leaf, or to the base of the panicle for flowering shoots. The basal diameter was taken in the middle of the second internode. Mean values per experimental plot were used for data analysis.

The aboveground biomass of all vascular plants except Phragmites (hereafter called "other species") was harvested in the last days of June and of August 1997, i.e. just before the mowing. To avoid confusion, the second sampling will again be called the "September" sampling. Block 2 could not be sampled for lack of time. The biomass was clipped at ground level in three 0.16-m² quadrats per plot. Mosses were not sampled because they were sparse at all sites. Care was taken in the control plots to sample new quadrats at the second date. The plant material was dried at 70 °C, weighed and ground. Total N and P were extracted using a modified Kjeldahl method (1h digestion at 420 °C with H₂SO₄ 98% and a copper sulphatetitane oxide catalyst). Concentrations of N and P were determined colorimetrically on a flow injection analyser (TECATOR, Höganäs, Sweden).

The biomass of *Phragmites* in the permanent quadrats was estimated non-destructively through field measurements and calibrations based on Güsewell & Klötzli (1997). For calibrations 20–40 shoots per experimental plot were harvested outside the permanent quadrats in late June 1996 and within the permanent quadrats in late August or early September 1996. The relationship between weight and length of individual shoots was

Table 1. Effect of an additional cutting in June on shoot density and size of Phragmites australis measured in late June and in early September from 1995 to 1997. Data are mean squares and the significance of effects based on repeated measures ANOVA. See text for details on calculations

Source	df	Late June			Early September		
		Density	Length	Diameter	Density	Length	Diameter
Treatment	1 3	149.30 ns	0.1 ns	0.03 ns	0.06 ns	0.25 ns	0.04 ns
Error		59.90	125.3	0.50	0.29	0.70	0.19
Year	1 3	166.57 < 0.01	1586.1 < 0.1	0.58 ns	0.70 < 0.1	1.30 < 0.05	0.01 ns
Error		4.20	202.1	0.22	0.10	0.05	0.00
Treat x year	1	69.97 ns	0.55 ns	0.01 ns	0.01 ns	1.09 ns	0.05 < 0.05
Error	3	29.08	8.22	0.06	0.04	0.28	0.00

used to estimate the biomass of *Phragmites* in 1996. The relationship between the mean shoot weight and the product of mean shoot length and mean basal diameter was used to estimate the biomass in 1995 and 1997 (cf. Güsewell 1997).

To determine nutrient concentrations in *Phragmites*, 20–40 shoots were harvested per experimental plot in the last days of June and August 1997 (outside the permanent quadrats in June) and analysed as described for the "other species".

DATA ANALYSIS

To analyse how the additional June cutting had influenced changes in the density and size of Phragmites shoots from 1995 to 1997, the 1995 data were used as a baseline, i.e. they were subtracted from the 1996 and 1997 data. The resulting differences were then analysed with univariate repeated measures ANOVA (factors treatment, df = 1, year, df = 1, and block, df = 3). Block 3 was not included in this analysis because mowing in June had only started in 1996. As significant "year x block" interactions occurred, "year" effects were tested against these interactions, and "treatment x year" interactions were tested against the residual error term. Aboveground biomass, nutrient concentrations, total nutrient contents and the contribution of *Phragmites* to the aboveground biomass were all log-transformed to stabilize error variances; they were analysed with two-way ANOVA with the factors "treatment" and "block", but only the "treatment" effect was tested (Sokal & Rohlf 1995, p. 347). Effects with type I error probabilities of 0.05–0.10 were considered to indicate tendencies not confirmed yet, but still relevant for conservation management.

Results

DENSITY AND SIZE OF PHRAGMITES

During the three years of the experiment, shoot density increased and shoot size decreased (Fig. 1), but this trend was independent of the additional mowing in June; differences between 1995 and the two subsequent years did not differ between treated plots (with additional June cutting) and control plots (Table 1). Shoot density in June, however, seemed to increase in treated plots compared with the controls (Fig. 1a). Thus, the additional mowing in June had certainly not reduced the abundance of *Phragmites*, and had possibly increased it.

Large differences in shoot size were found between treated and control plots in August (Fig. 1b) because shoots of different age were

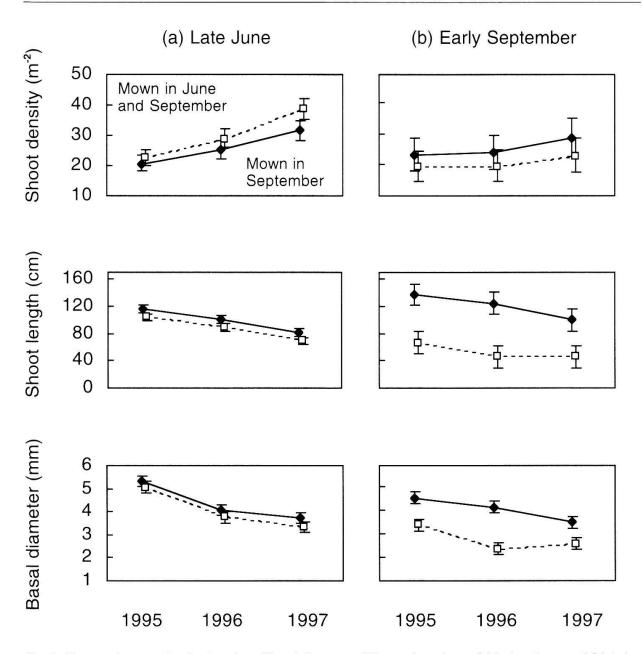


Fig. 1. Changes in mean density, length and basal diameter of Phragmites shoots (a) in late June, and (b) in late August–early September from 1995 (start of the experiment) to 1997. Dashed lines: plots mown in June and September. Solid lines: plots only mown in September. Error bars indicate standard errors of comparisons between treatments within each year (Mead 1988, p. 397).

compared: shoots had emerged in spring in the control plots, but only in July in the treated plots. However, these differences did not increase from 1995 to 1997. The significant "treatment x year" interaction for shoot diameter in August (Table 1) reflected a fluctuation, not an increasing treatment effect (cf. Fig. 1b). This means that the ability of *Phragmites* to regenerate after the June cutting

did not decrease in the treated plots during the investigation period.

ABOVEGROUND BIOMASS OF *PHRAGMITES* AND OF THE OTHER SPECIES

To determine how additional mowing in June affects the aboveground biomass, one can either consider regeneration after mowing, i.e. the biomass reached at the end of the same

Table 2. Effect of an additional cutting in June on aboveground biomass and total nutrient contents ("standing	ng
stock") of Phragmites australis, and all other species, measured in late June 1997.	

		Biomass	Nitrogen	Phosphorus
Phragmites				
Mean (g m ⁻²)	Treated plots Control plots	86.28 74.06	1.43 1.22	0.11 0.09
ANOVA results	$F_{t,3}$ Significance (P)	0.91 ns	1.68 ns	3.92 ns
Other species				
Mean (g m ⁻²)	Treated plots Control plots	245.25 313.88	3.71 4.80	0.30 0.38
ANOVA results	$F_{t,3}$ Significance (P)	15.67 <0.05	32.32 <0.05	19.52 <0.05

growing season, or the biomass produced during the following year(s).

Within the same growing season, mowing in June affected the aboveground biomass of *Phragmites* more strongly than that of the other species: the biomass of *Phragmites* at the end of the summer was only 37% (SE = 8%) of the biomass before mowing (end of June), compared with 61% (SE = 6%) for the other species. This difference could not be attributed to a seasonal effect, as the biomass of *Phragmites* and the biomass of the other species increased by the same factor (26–27%) from June to August in the control plots. Consequently, *Phragmites* regenerated poorly after mowing compared with other species.

Conversely, additional mowing in June during two years did not affect the above-ground biomass reached by *Phragmites* in June 1997 (third year), nor the amounts of nutrients stored in this biomass (Table 2a), whereas the biomass and nutrient contents of the other species were reduced (Table 2b). As a result, the contribution of *Phragmites* to the aboveground biomass in June 1997 was higher in treated plots than in the controls $(29\% \ vs.\ 21\%;\ F=7.03,\ P<0.1)$.

NUTRIENT ECONOMY OF *PHRAGMITES*AND THE OTHER SPECIES

Mowing in June during two years did not lead to lower nutrient concentrations in the aboveground biomass of treated plots compared with controls: neither for *Phragmites* nor for the other species did nutrient concentrations in June 1997 differ between treatments (Table 3a). Nutrient concentrations in early September were markedly higher in treated plots than in the controls, particularly for phosphorus (Table 3b), because in the treated plots shoots were much younger.

Mowing in June and September removed more nutrients than mowing only in September, even though the total biomass exported by mowing (adding together both harvests in the treated plots) was the same with both treatments (Table 4a). On an average, mowing in June and September removed 44% more N and 85% more P than mowing in September only. The greater nutrient export in treated plots was mainly due to the higher nutrient concentrations in September. Differences between treatments were stronger for *Phragmites* than for the other species: on an average, mowing in June and September re-

Table 3. Effects of an additional cutting in June on nutrient concentrations in the aboveground biomass of Phragmites australis, and all other species in the third year of the experiment

		N (June)	N (Sept)	P (June)	P (Sept)
Phragmites					and the second s
Mean (mg g ⁻¹)	Treated plots Control plots	16.76 16.54	17.36 11.69	1.32 1.26	1.51 0.64
ANOVA results	$F_{t,3}$ Significance (P)	0.02 ns	56.98 <0.01	1.27 ns	113.86 <0.01
Other species					100000000000000000000000000000000000000
Mean (mg g ⁻¹)	Treated plots Control plots	15.19 15.32	16.86 12.56	1.28 1.28	1.89 1.27
ANOVA results	$F_{t,3}$ Significance (P)	0.20 ns	28.65 <0.05	0.22 ns	45.38 <0.01

Table 4. Effects of an additional cutting in June on the export of biomass and nutrients through mowing, and the relative contribution of Phragmites australis to this export in the third year of the experiment

		Biomass	Nitrogen	Phosphorus
Total export		A		
Mean (g m ⁻²)	Treated plots Control plots	537.11 495.70	8.53 5.96	0.70 0.38
ANOVA results	$F_{I,3}$ Significance (P)	1.02 ns	20.83 <0.05	33.05 <0.05
Contribution of Phra	agmites			
Mean (%)	Treated plots Control plots	21.89 18.76	22.95 18.12	22.42 14.61
ANOVA results	$F_{t,3}$ Significance (P)	2.36 ns	2.83 ns	25.79 <0.05

moved 90% more N and 181% more P than mowing in September only for *Phragmites*, but only 30% more N and 64% more P for the other species. The contribution of *Phragmites* to the export of phosphorus was, therefore, higher in the treated plots than in the controls, and the same tendency (though not significant) was found for nitrogen (Table 4b). Thus, the nutrient economy of *Phragmites* appeared to be more strongly affected by an additional cutting in June than the nutrient economy of the other species.

Discussion

IMPACT OF MOWING ON THE NUTRIENT ECONOMY OF *PHRAGMITES*

In its main habitat, aquatic sites, *Phragmites australis* experiences virtually no interspecific competition, but in regularly managed fen meadows, there are many competitors, and *Phragmites* emerges later in the year than most of them (Hürlimann 1951; Buttery & Lambert 1965; Haslam 1971). To be successful under these conditions, *Phragmites* must

grow rapidly in early summer despite the shading and the root competition of species that have developed earlier. Carbohydrates and mineral nutrients stored in belowground parts during previous growing seasons make such a fast growth possible. Additional mowing in late June was expected to remove the carbohydrates and mineral nutrients that were supplied to the shoots in spring (Fiala 1976; Hocking 1989; Granéli 1990; Granéli et al. 1992), i.e. to deplete the belowground reserves, and thus, to reduce the competitive ability of *Phragmites* (e.g. Rodewald-Rudescu 1974; Egloff 1984).

In the present study, mowing in June and September did, indeed, remove significantly more nutrients than mowing only in September. The carbohydrate economy was not investigated, but in plots that had been mown in June, N and P concentrations were still as high in early September as in late June. This suggested that *Phragmites* was still growing, and that most assimilates produced during August were used for growth, whereas they would have been stored in the rhizomes if plots had not been mown in June (Fiala 1976; Granéli *et al.* 1992; Guthruf *et al.* 1993).

ABSENCE OF SHORT-TERM TREATMENT EFFECTS

In view of the impact of additional mowing in June on the nutrient economy of *Phragmites*, a negative effect on shoot size, aboveground biomass or nutrient contents could be expected. However, no such effect was observed during the period of investigation. Changes in shoot density and shoot size observed in treated plots between 1995 and 1997 did not differ from those observed in the controls. They were, therefore, unrelated to management, whereas differences in groundwater level and weather conditions (e.g. dry spring 1996, severe late frosts in April 1997)

probably had an impact (Haslam 1972). Since the biomass of the other species was, on an average, reduced by the additional mowing, the contribution of *Phragmites* to total aboveground biomass was even increased, suggesting that the species had actually been promoted by this treatment.

The most obvious explanation for the absence of significant treatment effects was the limited duration of the experiment. In long-term management experiments, time lags of up to ten years were observed between a change in management and changes in the abundance of certain species (Olff & Bakker 1991; Grootjans *et al.* 1996). Rhizomatous geophytes like *Phragmites* seem particularly able to respond with a delay because they only use a fraction of their belowground reserves for the annual growth (Granéli *et al.* 1992), and therefore stores will take several years to be depleted.

However, mowing did significantly reduce the abundance of *Phragmites* within one or two years in other experiments. There are several possible reasons for the stronger treatment effects observed in those studies:

- Some of the experiments were carried out at unmanaged sites (Gryseels 1989; Briemle & Ellenberg 1994), where the accumulated litter of *Phragmites* probably excluded other species (Haslam 1971; George 1992). Litter removal allowed other species to invade and to compete against *Phragmites*.
- Some experiments were conducted on waterlogged soils. Removing the culms reduces the supply of oxygen to the belowground parts (Brix 1990; Armstrong & Armstrong 1988). The subsequent anoxia causes important losses of carbohydrates from rhizomes (Cízková-Koncalová et al. 1992), which may strongly affect the plant (Weisner & Granéli 1989).

- In aquatic stands, stubbles were flooded after mowing, and rhizomes began to rot (Husák 1978); salt water proved to be particularly detrimental (Hellings & Gallagher 1992).
- In highly productive terrestrial stands, *Phragmites* was displaced by certain grass species, e.g. *Agrostis stolonifera* or *Glyceria maxima*, which are more tolerant to mowing in summer than *Phragmites* (George 1992; Rodwell 1995).

UNCERTAIN LONG-TERM TREATMENT EFFECTS

Assuming that *Phragmites* would strongly decrease once its belowground reserves are depleted, it would be interesting to estimate the time needed for this depletion. Unfortunately, this is hardly possible without field measurements of belowground biomass and its turnover. Direct measurements at the experimental sites are needed because the belowground biomass of *Phragmites* as well as its production and decay vary considerably among sites (Fiala 1976; Hocking 1989; Granéli et al. 1992; Cízková et al. 1996). Due to the depth of the rhizomes and roots of *Phragmites* at terrestrial sites and to their spatial variability (Kvet 1973; Ondok 1978), such measurements would have been both unreliable and destructive and were omitted in this study.

Moreover, internal nutrient cycling is not particular to *Phragmites*, but typical for the dominant species in fen meadows (Kuhn *et al.* 1982; Ganzert & Pfadenhauer 1986; Bernard *et al.* 1988; Marti 1994). Consequently, other species are likely to be affected by nutrient depletion as much as *Phragmites*. In this experiment, additional mowing in June removed more nutrients from *Phragmites* than from the other species. However, this result is probably mainly due to the fact that, in opposition to *Phragmites*, the other species had a

lower June biomass in the treated plots than in the controls. Without this difference in June biomass, the additional nutrient removal through the mowing in June would have been even lower for *Phragmites* than for the other species because *Phragmites* regenerated poorly after the first cut.

The long-term effect of mowing is likely to depend on the nutrient status of sites. Experiments by Rosenthal (1992) show that even mowing twice a year may be insufficient to reduce invasive rhizomatous species in nutrient-rich fen meadows. However, according to Briemle & Ellenberg (1994), *Phragmites* is more sensitive to moving than the species investigated by Rosenthal (1992). It might, therefore, decrease in plots mown in June for a longer time. The vertical distribution of nutrients in the soil is also important: an inflow of nutrient-rich groundwater at a depth within reach of Phragmites roots, but not of roots of smaller species (Boller-Elmer 1977; Klötzli 1986), might enable Phragmites to take up more nutrients than other species. It might then retain the initial advantage apparent in this study.

CONSEQUENCES FOR MANAGEMENT

To evaluate whether mowing in June and September or another form of management is more suitable for the conservation or restoration of reed-invaded fen meadows, effects of other possible treatments on *Phragmites*, on the other species and on site productivity need to be considered.

The control treatment in the present study, i.e. mowing once a year in September, is likely to be milder than mowing twice a year and seems therefore preferable for the typical plant species of fen meadows (cf. Briemle & Ellenberg 1994). However, this treatment exports relatively few nutrients because aboveground biomass in late summer repre-

sents only part of the annual production (e.g. Bernard et al. 1988), and because translocation of nutrients to belowground parts already starts in July or August in many dominant plant species (e.g. Warnke-Grüttner 1990). Moreover, the total nutrient contents (biomass x nutrient concentrations) in the control plots of the present study indicated that Phragmites had already translocated a greater fraction of nutrients to rhizomes than the other species before the September cutting. This is consistent with results of Warnke-Grüttner (1990) and suggests that the management currently practised in most fen meadows is rather unsuitable if the aim is to prevent a spread of *Phragmites* (Hürlimann 1951; Rosenthal 1992; Schütz & Ochse 1997).

More frequent mowing, i.e. three or four times a year, might be more effective in reducing dominant species like *Phragmites* (Rosenthal 1992) and in lowering site productivity than mowing twice a year. This kind of management would probably lead to undesirable changes in species composition, i.e. promote species adapted to frequent mowing (Kapfer & Pfadenhauer 1986). Kapfer (1987) and Klötzli (1991) recommended to only apply it if biomass production exceeds 500 g m⁻² or if no rare or characteristic fen species are present at a site.

Several studies have shown that grazing can be a very effective measure to reduce *Phragmites* because this species is selectively eaten by cattle or horses and severely affected by trampling (van Deursen & Drost 1990; Roze 1993; Walther 1994). This management has now also been implemented in certain Swiss fens, using highland cattle (Hasler 1996a,b). One important drawback of grazing is that it removes much less nutrients than mowing (Bakker 1989; Marrs 1993). Moreover, other species (e.g. orchids) are likely to suffer from trampling as well. A combination of mowing

and short periods of grazing when the soil is dry might be a suitable, but rather labourintensive solution.

Since each of the possible alternative treatments presents certain drawbacks, mowing in June and September might still be a suitable management for reed-invaded fen meadows, even if short-time success cannot be expected and long-term effects still need to be evaluated. At least, this mowing regime leads to a considerably higher nutrient export than mowing in September only, to a lower aboveground biomass and a more open vegetation structure during most of the summer. These effects are probably more important for species richness and for the rare species of fen meadows than a possible (future) reduction of *Phragmites* (Güsewell & Klötzli 1998).

Acknowledgements

I am very grateful to the conservation authorities (Amt für Raumplanung Kanton Zürich) and to the landowners for allowing experiments on private land in nature reserves, to S. Bertschinger, T. Fotsch and T. Weber for mowing the experimental plots and to U. Bollens, A. Brülisauer, A. Gerster, A. Hegi, M. Schaub and R. Trachsler for kind assistance and enjoyable company in field and laboratory work. A. Bosshard, F. Klötzli, J. Kollmann, D. Matthies, D. Ramseier, B. Roy and an anonymous referee gave valuable comments on successive drafts of the manuscript. The research was supported by a grant from the Swiss Federal Office for Environment, Forest and Landscape (BUWAL).

References

Armstrong J. & Armstrong W. (1988) *Phragmites* australis – a preliminary study of soil oxidizing sites and internal gas transport pathways. *New Phytologist*, **108**, 373–382.

- Bakker, J.P. (1989) Nature management by grazing and cutting. *Geobotany*, 14, 1–400.
- Bakker, J.P. & de Vries, Y. (1985) Über die Wiederherstellung artenreicher Wiesengesellschaften unter verschiedenen Mahdsystemen in den Niederlanden. *Natur und Landschaft*, **60**, 292–296.
- Bernard, J.M., Solander, D. & Kvet, J. (1988) Production and nutrient dynamics in *Carex* wetlands. *Aquatic Botany*, **39**, 125–147.
- Bibby, C.J. & Lunn, J. (1982) Conservation of reedbeds and their avifauna in England and Wales. *Biological Conservation*, 23, 167–186.
- Biewer, H. (1994) Landschaftsökologisch-moorkundliche Untersuchungen im Osterried bei Laupheim. II. Vegetation und Flora. *Feuchtgebiete: Gefährdung, Schutz, Renaturierung* (eds. R. Böcker & A. Kohler), 271–276. Verlag Günter Heimbach, Ostfildern.
- Boller-Elmer, K. (1977) Stickstoff-Düngungseinflüsse von Intensiv-Grünland auf Streu- und Moorwiesen. Veröffentlichungen des Geobotanischen Instituts ETH, Stiftung Rübel, Zürich, 63, 1–103.
- Bressous, B., Broggi, M.F., Hintermann, U., Grünig, A., Marti, K., Schlegel, H. & Theis, E. (1992) Anforderungen beim Schutz der Flachmoore von nationaler Bedeutung. Handbuch Moorschutz in der Schweiz Band II, Beitrag 1.1.2. EDI, BUWAL, Bern.
- Briemle, G. & Ellenberg, H. (1994) Zur Mahdverträglichkeit von Grünlandpflanzen. *Natur und Landschaft*, **69**, 139–147.
- Brix, H. (1990) Gas exchange through the soil-atmosphere interphase and through dead culms of *Phragmites australis* in a constructed reed bed receiving domestic sewage. *Water Research*, 24, 259–266.
- Buttery, B.R. & Lambert, J.M. (1965) Competition between *Glyceria maxima* and *Phragmites communis* in the region of Surlingham Broad. I. The competition mechanism. *Journal of Ecology*, 53, 163–181.
- Cízková-Koncalová, H.J., Kvet, J. & Thompson, K. (1992) Carbon starvation: a clue to reed decline in eutrophicated lakes. *Aquatic Botany*, 43, 105–113.
- Cízková, H., Lukavská, J., Priban, K., Kopecky, J. & Brabcová, H. (1996) Carbohydrate levels in rhizomes of *Phragmites australis* at an oligotrophic and a eutrophic site: a preliminary study.

- Folia Geobotanica et Phytotaxonomica, 31, 111–118.
- Demina, O.M. (1979) Changes in reed meadows when mown for hay in the lower reaches of southern Kazakhstan rivers. *Botanicheskii Zhurnal*, **64**, 58–64.
- Egloff, T. (1984) Richtlinien zur Bewirtschaftung und Pflege von Riedern und Mooren im Sinne des Naturschutzes. SBN, Basel.
- Egloff, T. (1986) Auswirkungen und Beseitigung von Düngungseinflüssen auf Streuwiesen. Eutrophierungssimulation und Regenerationsexperimente im nördlichen Schweizer Mittelland. Veröffentlichungen des Geobotanischen Instituts ETH, Stiftung Rübel, Zürich, 89, 1–183.
- Fiala, K. (1976) Underground organs of *Phragmites* communis, their growth, biomass and net production. *Folia Geobotanica et Phytotaxonomica*, 11, 225–259.
- Finckh, B. (1960) Umbruchlose Verbesserung ertragsarmer Streuwiesen. *Bayerisches Landwirtschaftliches Jahresbuch*, 37, 97-119.
- Ganzert, C. & Pfadenhauer, J. (1986) Seasonal dynamics of shoot nutrients in *Schoenus ferrugineus* (Cyperaceae). *Holarctic Ecology*, **9**, 137–142.
- George, M. (1992) The Land Use, Ecology and Conservation of Broadland. Packard Publishing, Chichester.
- Granéli, W. (1984) Reed *Phragmites australis* (Cav.) Trin. ex Steudel as an energy source in Sweden. *Biomass*, 4, 183–208.
- Granéli, W. (1990) Standing crop and mineral content of reed in Sweden management of reed stands to maximize harvestable biomass. *Folia Geobotanica et Phytotaxonomica*, **25**, 291–302.
- Granéli, W., Weisner, S.E.B. & Sytsma, M.D. (1992) Rhizome dynamics and resource storage in Phragmites australis. Wetlands Ecology and Management, 1, 239–247.
- Grootjans, A.P., Fresco, L.F.M., de Leeuw, C.C. & Schipper, P.C. (1996) Degeneration of speciesrich *Calthion palustris* hay meadows: some considerations on the community concept. *Journal of Vegetation Science*, 7, 185–194.
- Gryseels, M. (1989) Nature management experiments in a derelict reedmarsh. II. Effects of summer mowing. *Biological Conservation*, 48, 85–99.
- Güsewell, S. (1997) Evaluation and management of fen meadows invaded by Phragmites australis. PhD thesis, ETH, Zürich.

- Güsewell, S. & Klötzli, F. (1997) Measuring the abundance of *Phragmites communis* Trin. in wet meadows: A methodological investigation. *Bulletin of the Geobotanical Institute ETH*, **63**, 11–24.
- Güsewell, S. & Klötzli, F. (1998) Abundance of common reed (*Phragmites australis*), site conditions and conservation value of fen meadows. *Acta Botanica Neerlandica*, 47, 113–129.
- Guthruf, K., Zenger, C. & Brändle, R. (1993) The habitat dependent productivity of reed (*Phragmites australis*) and its significance. *Seeuferzerstörung und Seeuferrenaturierung in Mitteleuropa* (eds. W. Ostendorp & P. Krumscheid-Plankert), pp. 1–8. Gustav Fischer, Stuttgart.
- Haslam, S.M. (1971) Community regulation in *Phragmites communis*. II. Mixed stands. *Journal of Ecology*, **59**, 75–88.
- Haslam, S.M. (1972) Biological flora of the British Isles. *Phragmites communis* Trin. *Journal of Ecology*, **60**, 585–610.
- Haslam, S.M. (1973a) The management of British wetlands. I. Economic and amenity use. *Journal of Environmental Management*, 1, 303–320.
- Haslam, S.M. (1973b) The management of British wetlands. II. Conservation. *Journal of Environmental Management*, 1, 345–361.
- Hasler, A. (1996a) Extensive Beweidung als neue Pflegemassnahme in Riedgebieten? *Naturschutz in der Gemeinde*, **12/1996**, 11–12.
- Hasler, A. (1996b) Mit Rindern gegen Büsche und Schilf. *Ornis*, 4/96, 14–15.
- Hellings, S.E. & Gallagher, J.L. (1992) The effects of salinity and flooding in *Phragmites australis*. *Journal of Applied Ecology*, **29**, 41–49.
- Hess, H.E., Landolt, E. & Hirzel, R. (1991) *Bestimmungsschlüssel zur Flora der Schweiz*. Birkhäuser, Basel.
- Hocking, P.J. (1989) Seasonal dynamics of production, and nutrient accumulation and cycling by *Phragmites australis* (Cav.) Trin. ex Steudel in a nutrient-enriched swamp in inland Australia. I. Whole plants. *Australian Journal of Marine and Freshwater Research*, 40, 421–444.
- Hürlimann, H. (1951) Zur Lebensgeschichte des Schilfs an den Ufern der Schweizer Seen. Beiträge zur Geobotanischen Landesaufnahme der Schweiz, 30, Bern.
- Husák, S. (1978) Control of reed and reed mace stands by cutting. *Ecological Studies*, **28**, 404–408.

- Izatt, L. (1979) *Phragmites* a real reed problem. *Cane growers' Quarterly Bulletin*, **43**, 40–41.
- Jones, W.L. & Lehman, W.C. (1987) *Phragmites* control and revegetation following aerial applications of Glyphosate in Delaware. Waterfowl & Wetlands Symposium. *Proceedings of a Symposium on Waterfowl and Wetlands Management in the Coastal Zone of the Atlantic Flyway.* (eds. W.R. Whitman & W.H. Meredith), pp. 184–196. Delaware Department of natural Resources and Environmental Control, Dover, Delaware.
- Kapfer, A. (1987) Untersuchungen zur Renaturierung ehemaliger Streuwiesen im südwestdeutschen Alpenvorland erste Tendenzen nach 4 Versuchsjahren. Erfassung und Bewertung anthropogener Vegetationsveränderungen (eds. R. Schubert & W. Hilbig), pp. 179–215. Martin-Luther-Universität, Halle-Wittenberg.
- Kapfer, A. & Pfadenhauer, J. (1986) Vegetationskundliche Untersuchungen zur Pflege von Pfeifengras-Streuwiesen. Natur und Landschaft, 61, 428–432.
- Klötzli, F. (1974) Über Belastbarkeit und Produktion in Schilfröhrichten. *Verhandlungen der Gesellschaft für Ökologie*, 2, 237–247.
- Klötzli, F. (1986) Tendenzen zur Eutrophierung in Feuchtgebieten. Veröffentlichungen des Geobotanischen Instituts ETH, Stiftung Rübel, Zürich, 87, 343–361.
- Klötzli, F. (1991) Möglichkeiten und erste Ergebnisse mitteleuropäischer Renaturierungen. *Verhandlungen der Gesellschaft für Ökologie*, **20**, 229–242.
- Kuhn, M.K., Leupi, E., Oberholzer-Tschütscher, B. & Oberholzer, H.R. (1982) Rückzug von Nährstoffen bei *Molinia caerulea* im Herbst. *Berichte des Geobotanischen Instituts ETH*, Stiftung Rübel, Zürich, 49, 146–153.
- Kvet, J. (1973) Mineral nutrients in shoots of reed (*Phragmites communis*). *Polskie Archiwum Hydrobiologii*, **20**, 137–147.
- Marrs, R.H. (1993) Soil fertility and nature conservation in Europe. Theoretical considerations and practical management solutions. *Advances in Ecological Research*, 24, 241–300.
- Marti, K. (1994) Zum Standort von *Magnocaricion*-Gesellschaften in der Schweiz. *Veröffentlichungen des Geobotanischen Instituts ETH, Stiftung Rübel, Zürich,* 120, 1–97.
- Marti, K. & Müller, R. (1993) Überprüfung des Pufferzonen-Schlüssels der BFÖ (1986) auf seine

- Wirksamkeit in der Praxis. Bericht zuhanden des Amtes für Raumplanung des Kantons Zürich, Zürich.
- Mead, R. (1988) *The Design of Experiments*. Cambridge University Press, Cambridge.
- Olff, H. & Bakker, J.P. (1991) Long-term dynamics of standing crop and species composition after the cessation of fertilizer application to mown grassland. *Journal of Applied Ecology*, **28**, 1040–1052.
- Ondok, J.P. (1978) Estimation of seasonal growth of underground biomass. *Pond Littoral Ecosystems. Ecological Studies 28* (eds. D. Dykyjová & J. Kvet), pp. 193–197. Springer, Berlin.
- Oomes, M.J.M. & Altena, H.J. (1987) Changes in the vegetation of extensively used agricultural grassland caused by cutting date and cutting frequency. *Erfassung und Bewertung anthropogener Vegetationsveränderungen* (eds. R. Schubert & W. Hilbig), pp. 75–91. Martin-Luther-Universität, Halle-Wittenberg.
- Ostendorp, W. (1995) Effect of management on the mechanical resistance of lakeside reeds in Lake Constance. *Acta Oecologica*, **16**, 277–294.
- Rodewald-Rudescu, L. (1974) Das Schilfrohr (Phragmites communis *Trinius*). Schweizerbart'sche Verlagsbuchhandlung, Stuttgart.
- Rodwell, J.S. (1995) *Aquatic Communities, Swamps* and *Tall-herb Fens*. British Plant Communities, Cambridge.
- Rosenthal, G. (1992) Problempflanzen bei der Extensivierung von Feuchtwiesen. *NNA-Berichte*, 5, 27–36.
- Rowell, T.A., Guaring, L. & Harvey, H.J. (1985) The experimental management of vegetation at Wicken Fen, Cambridgeshire. *Journal of Applied Ecology*, **22**, 217–227.
- Roze, F. (1993) Successions végétales après pâturage extensif par des chevaux dans une roselière. *Bulletin d'Ecologie*, 24, 203–209.
- Schütz, P. & Ochse, M. (1997) Effizienzkontrolle von Pflege- und Entwicklungsplänen für Schutzgebiete in Nordrhein-Westfalen. *Naturschutz und Landschaftsplanung*, 29, 20–31.
- Sokal, R.R. & Rohlf, J. (1995) *Biometry*. Freeman & Company, New York.
- Tscharntke, T. (1992) Fragmentation of *Phragmites* habitats, minimum viable population size, habitat suitability, and local extinction of moths, midges, flies, aphids and birds. *Conservation Biology*, **6**, 530–536.

- Van Deursen, E.J.M. & Drost, H.J. (1990) Defoliation and treading by cattle of reed *Phragmites australis*. *Journal of Applied Ecology*, 27, 284–297.
- Van Diggelen, R., Molenaar, W. J. & Kooijman, A.M. (1996) Vegetation succession in a floating mire in relation to management and hydrology. *Journal of Vegetation Science*, 7, 809–820.
- Walther, B. (1994) *Biomanagement mit dem Schottischen Hochlandrind*. PhD thesis, University Basel.
- Warnke-Grüttner, R. (1990) Ökologische Untersuchungen zum Nährstoff- und Wasserhaushalt in Niedermooren des Westlichen Bodenseegebiets. *Dissertationes Botanicae*, 148, 1–214.
- Weisner, S.E.B. & Granéli, W. (1989) Influence of substrat conditions on the growth of *Phragmites australis* after a reduction in oxygen transport to below-ground parts. *Aquatic Botany*, **35**, 71–80.
- Wolf, G., Wiechmann, H. & Forth, K. (1984) Vegetationsentwicklung in aufgegebenen Feuchtwiesen und Auswirkungen von Pflegemassnahmen auf Pflanzenbestand und Boden. *Natur und Landschaft*, 59, 316–322.

Received 2 February 1998 revised version accepted 1 April 1998

8

,